

## APPLIED ECOLOGY

## Environmental controls on modern scleractinian coral and reef-scale calcification

Travis A. Courtney,<sup>1\*</sup> Mario Lebrato,<sup>1,2</sup> Nicholas R. Bates,<sup>3,4</sup> Andrew Collins,<sup>3</sup> Samantha J. de Putron,<sup>3</sup> Rebecca Garley,<sup>3</sup> Rod Johnson,<sup>3</sup> Juan-Carlos Molinero,<sup>5</sup> Timothy J. Noyes,<sup>3</sup> Christopher L. Sabine,<sup>6</sup> Andreas J. Andersson<sup>1</sup>

Modern reef-building corals sustain a wide range of ecosystem services because of their ability to build calcium carbonate reef systems. The influence of environmental variables on coral calcification rates has been extensively studied, but our understanding of their relative importance is limited by the absence of in situ observations and the ability to decouple the interactions between different properties. We show that temperature is the primary driver of coral colony (*Porites astreoides* and *Diploria labyrinthiformis*) and reef-scale calcification rates over a 2-year monitoring period from the Bermuda coral reef. On the basis of multimodel climate simulations (Coupled Model Intercomparison Project Phase 5) and assuming sufficient coral nutrition, our results suggest that *P. astreoides* and *D. labyrinthiformis* coral calcification rates in Bermuda could increase throughout the 21st century as a result of gradual warming predicted under a minimum CO<sub>2</sub> emissions pathway [representative concentration pathway (RCP) 2.6] with positive 21st-century calcification rates potentially maintained under a reduced CO<sub>2</sub> emissions pathway (RCP 4.5). These results highlight the potential benefits of rapid reductions in global anthropogenic CO<sub>2</sub> emissions for 21st-century Bermuda coral reefs and the ecosystem services they provide.

## INTRODUCTION

Tropical coral reef ecosystems provide humanity with a range of direct (tourism and fishing), indirect (shoreline protection and fisheries recruitment), and nonuse (biodiversity and intrinsic value) ecosystem services (1). The structure of these ecosystems is maintained by net positive production of calcium carbonate (CaCO<sub>3</sub>), with scleractinian corals accounting for the majority of the total coral reef CaCO<sub>3</sub> production (2). Current estimates suggest that there are fewer than 1000 zooxanthellate hermatypic scleractinian coral species sustaining the structural habitat that covers just 0.2% of the total ocean surface area and yet supports an estimated 35% of all species living in the global oceans (3).

The geologic record shows the onset of large and widespread coral reef accretion dating back to the late Triassic (~230 million years ago), when scleractinian corals are hypothesized to have first acquired photosynthetic zooxanthellae symbionts (4). This symbiosis allowed corals to expand geographic ranges (4) due to the algal symbionts providing upward of 100% of the daily respiratory carbon needed to sustain the modern coral host (5). The efficiency of this symbiosis allows corals to maintain primary productivity rates (8.0 to 40.0 g C m<sup>-2</sup> day<sup>-1</sup>) that are orders of magnitude higher than adjacent ocean water primary productivity (0.01 to 0.65 g C m<sup>-2</sup> day<sup>-1</sup>) (5).

Energy requirements for calcification initially suggest that photosynthesis is a key driver of coral calcification rates (5); however, coral calcification depends on a broad range of environmental variables, including seawater temperature, seawater carbonate chemistry, light and depth, food availability, nutrients, water flow rates, sedimentation, and competition (2). Although laboratory experiments have successfully

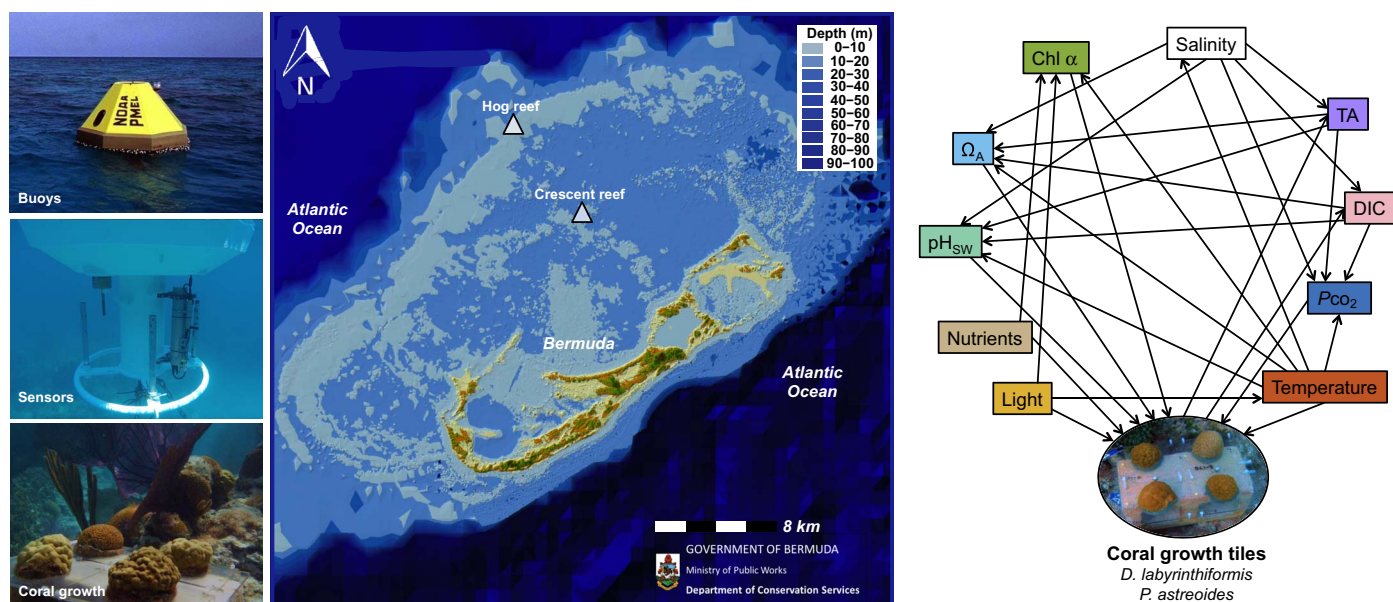
established relationships between coral calcification rates and independently altered environmental parameters (for example, temperature, light, pH, and seawater saturation state with respect to aragonite,  $\Omega_A = [\text{Ca}^{2+}][\text{CO}_3^{2-}]/K_{\text{sp}}$ ), comparatively fewer studies have explored the combined effects and/or the relative importance of these parameters under controlled laboratory conditions (6, 7) or under naturally variable in situ conditions (2, 8–11). The ability to measure and establish the relative importance of different drivers of calcification in the field is limited by the capacity to adequately monitor relevant properties simultaneously for extended periods of time (12) and to decouple the range of highly correlated and interdependent interactions between environmental factors (Fig. 1) (8). For example, temperature not only directly influences coral calcification rates but also strongly controls seawater pH and  $\Omega_A$ , which are both hypothesized to be independently important drivers of coral calcification (Fig. 1) (13). Temperature is also directly related to light availability and season, which affect coral calcification directly via light-enhanced calcification (Fig. 1) (2, 10) and indirectly via increased food availability resulting from seasonal patterns in oceanic primary production (Fig. 1) (2, 14). Characterizing the mechanisms and the relative importance of different drivers of coral calcification rates both in the current natural environment and under future ocean warming and acidification is essential for understanding how coral reefs calcify at present and for predicting how calcification rates and coral reef accretion will change under predicted near-future conditions.

The Bermuda coral reef is located at the northern limit of coral reefs in the North Atlantic Ocean (Fig. 1). Because of its relatively high latitude (32°N), Bermuda experiences greater seasonal differences and variations in environmental parameters than reefs located closer to the equator (15). Therefore, the Bermuda coral reef provides an excellent natural laboratory to explore the relationships between coral calcification rates and multiple environmental parameters. Over a 2-year period (August 2010 to September 2012), in situ environmental properties {temperature, seawater carbonate chemistry [dissolved inorganic carbon (DIC), total alkalinity (TA), partial pressure of CO<sub>2</sub> ( $P_{\text{CO}_2}$ ), pH<sub>sw</sub>,  $\Omega_A$ ], light, chlorophyll *a*, and inorganic nutrients} were characterized alongside coral calcification rates by two scleractinian reef-building

Copyright © 2017  
The Authors, some  
rights reserved;  
exclusive licensee  
American Association  
for the Advancement  
of Science. No claim to  
original U.S. Government  
Works. Distributed  
under a Creative  
Commons Attribution  
NonCommercial  
License 4.0 (CC BY-NC).

<sup>1</sup>Scripps Institution of Oceanography, University of California, San Diego, La Jolla, CA 92093, USA. <sup>2</sup>Christian-Albrechts-University Kiel, Kiel, Germany. <sup>3</sup>Bermuda Institute of Ocean Sciences, St. George's, Bermuda. <sup>4</sup>Department of Ocean and Earth Science, National Oceanography Centre Southampton, University of Southampton, Southampton, UK. <sup>5</sup>GEOMAR Helmholtz Center for Ocean Research, Marine Ecology/Food Webs, Kiel, Germany. <sup>6</sup>Pacific Marine Environmental Laboratory, National Oceanic and Atmospheric Administration, Seattle, WA 98115, USA.

\*Corresponding author. Email: traviscourtney@gmail.com



**Fig. 1. Map of Bermuda study sites and environmental controls on calcification.** The buoy, sensors, and in situ growth are presented next to a bathymetry map of Bermuda showing the locations of Hog Reef and Crescent Reef. Structural equation modeling (SEM) connections show the interactions between environmental drivers and their effect on coral calcification. Chl  $\alpha$ , chlorophyll  $\alpha$ . (The Bermuda Map is courtesy of M. Shailer of the Department of Conservation Services, Government of Bermuda.)

corals, *Porites astreoides* and *Diploria labyrinthiformis*, at two distinct reef environments. Colonies of *P. astreoides* and *D. labyrinthiformis* were transplanted onto tiles at a rim reef environment, Hog Reef, and an outer lagoon reef environment, Crescent Reef ( $n = 24$  colonies per species per site; Fig. 1), and buoyant-weighted every 2 to 3 months to measure in situ calcification rates (16). In addition, net ecosystem calcification (NEC = gross calcification – gross  $\text{CaCO}_3$  dissolution) for Hog Reef was calculated based on alkalinity anomalies (17) using monthly surface seawater TA samples and monthly offshore surface seawater TA from the Bermuda Atlantic Time-series Study (BATS) site (18) combined with estimates of seawater residence time (19). Assuming that calcification and  $\text{CaCO}_3$  dissolution are the only processes significantly influencing salinity-normalized seawater TA, this variable serves as a direct proxy of NEC and changes in a ratio of 2:1 for every net mole of  $\text{CaCO}_3$  deposited (17).

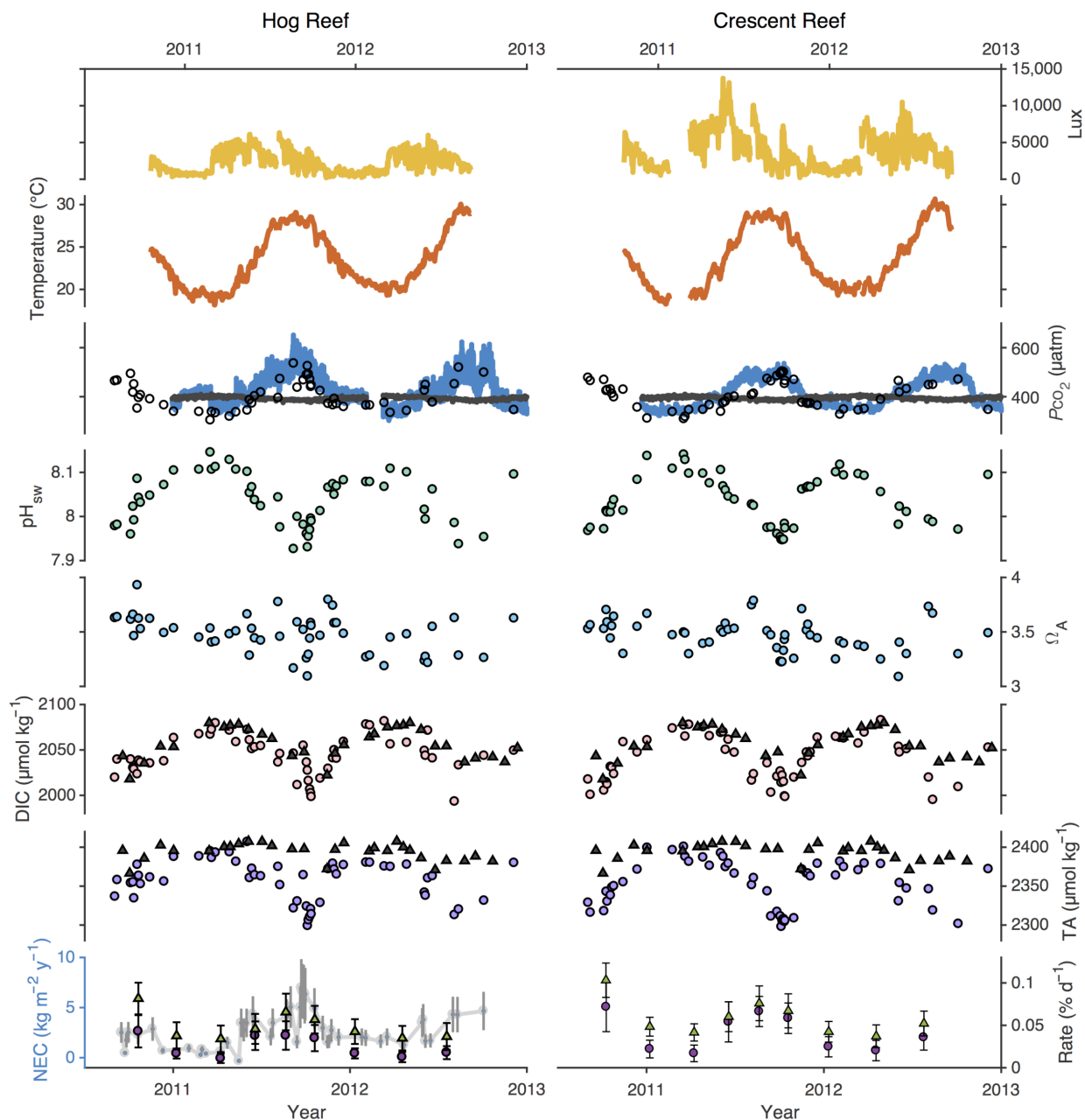
## RESULTS

### Observations of environmental controls and calcification

The data reveal a strong seasonal pattern in environmental variables and coral calcification rates over the 2-year study period (Fig. 2 and fig. S1). Light varied seasonally, with the lowest light intensity around January and the highest light intensity around June to July. Temperature lagged light intensity by 1 to 2 months, and maximum seawater temperatures were observed between August and October (Crescent Reef maximum temperature =  $30.7^\circ \pm 0.1^\circ\text{C}$ ), whereas minimum seawater temperatures occurred from January to March (Hog Reef minimum temperature =  $18.2^\circ \pm 0.1^\circ\text{C}$ ). Seasonal variability in seawater temperature and light was greater for Crescent Reef ( $\Delta$  temperature =  $12.4^\circ\text{C}$ ,  $\Delta$  light = 13,500 lux) than for Hog Reef ( $\Delta$  temperature =  $11.9^\circ\text{C}$ ,  $\Delta$  light = 6100 lux). Surface seawater  $P_{\text{CO}_2}$  was highest in the summer (Hog Reef maximum  $P_{\text{CO}_2} = 652 \mu\text{atm}$ ), well exceeding equilibrium with the atmosphere, whereas  $\text{pH}_{\text{sw}}$  was lowest in the summer (Hog Reef minimum  $\text{pH}_{\text{sw}} = 7.93$ ). The opposite trends were observed during winter (Hog Reef minimum  $P_{\text{CO}_2} = 303 \mu\text{atm}$ , maximum  $\text{pH}_{\text{sw}} = 8.14$ ).

These observed seasonal variations in  $P_{\text{CO}_2}$  and  $\text{pH}_{\text{sw}}$  can mainly be explained by the seasonal variability in temperature (for example, warming explains  $96 \pm 3\%$  of  $\Delta P_{\text{CO}_2}$  and  $90 \pm 3\%$  of  $\Delta \text{pH}_{\text{sw}}$  observed between 8 September 2010 and 25 February 2011). Seawater  $\Omega_A$  did not follow a strong seasonal variation similar to  $P_{\text{CO}_2}$  and  $\text{pH}_{\text{sw}}$ , although minimum and maximum values were observed in the winter and summer, respectively, ranging from 3.09 to 3.93 at Hog Reef. Reef seawater DIC and TA were strongly depleted in the summer relative to offshore, reflecting the uptake of DIC and calcium used for reef-scale net organic carbon production and net calcification. Seasonal variability in seawater carbonate chemistry was greater for Hog Reef ( $\Delta P_{\text{CO}_2} = 349 \mu\text{atm}$ ,  $\Delta \text{pH}_{\text{sw}} = 0.22$ ,  $\Delta \Omega_A = 0.84$ ) than for Crescent Reef ( $\Delta P_{\text{CO}_2} = 213 \mu\text{atm}$ ,  $\Delta \text{pH}_{\text{sw}} = 0.19$ ,  $\Delta \Omega_A = 0.70$ ), primarily due to higher biomass relative to water volume driving greater reef metabolic effects on the water column at Hog Reef (that is, mean  $\pm$  SE hard coral cover of  $28 \pm 1\%$  versus  $13 \pm 1\%$  at Crescent Reef) (20).

Notably, independently measured colony weight-normalized calcification rates of *P. astreoides* and *D. labyrinthiformis* colonies followed a similar seasonal variability with maximum rates observed from August to November and minimum rates observed from February to April. All forms of inorganic nitrogen, silica, and phosphorus remained low throughout the year but with occasional pulses (fig. S1). Monthly satellite chlorophyll  $a$  did not show strong seasonal variability at Hog Reef and Crescent Reef (fig. S1). Thus, maximum rates of both coral and reef-scale calcification appear to lag light intensity but coincide with maxima of seawater temperature,  $P_{\text{CO}_2}$ , and  $\Omega_A$  and minimum  $\text{pH}_{\text{sw}}$ . The seemingly paradoxical calcification maxima under seawater  $P_{\text{CO}_2}$  maximum and  $\text{pH}_{\text{sw}}$  minimum contradict the traditional understanding of the effects of seawater acidification on calcification (13), including a previous preliminary study conducted in the same area (21), but agree with recent studies that show that calcification can increase up to a threshold  $P_{\text{CO}_2}$  and pH if corals are adequately fed (6, 7) or if  $P_{\text{CO}_2}$ -pH conditions are within the natural variability experienced by the coral (22).

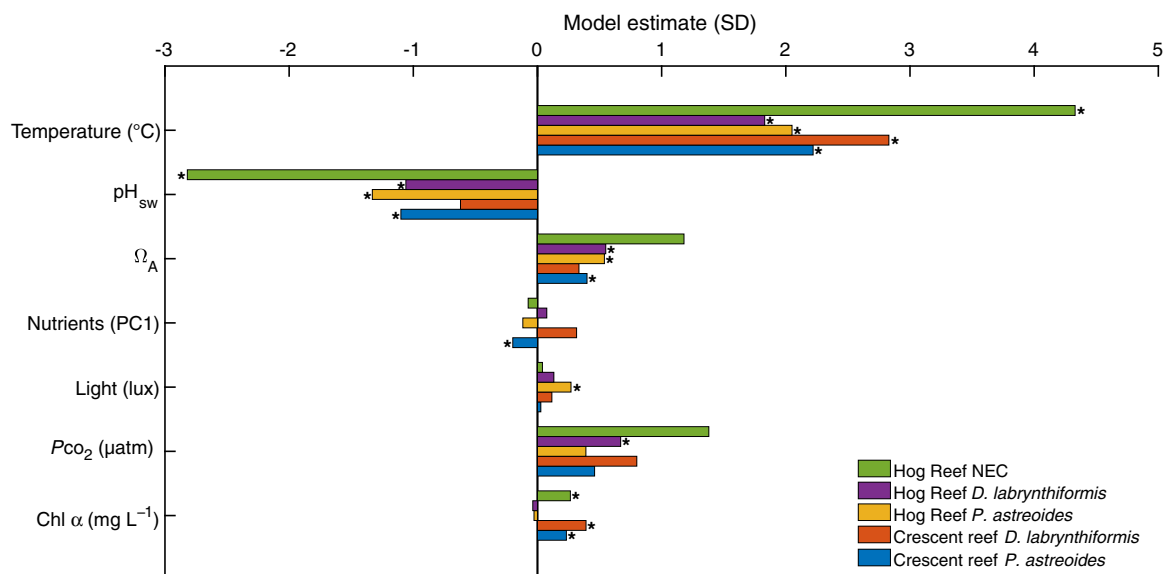


**Fig. 2. Hog Reef and Crescent Reef environmental data and coral calcification rates.** Both environmental parameters and calcification reveal strong seasonal trends in this high-latitude (32°N) coral reef system. For  $P_{CO_2}$  plots, blue denotes measured seawater  $P_{CO_2}$ , gray indicates measured atmospheric  $P_{CO_2}$ , and open circles represent seawater  $P_{CO_2}$  calculated from DIC and TA bottle samples. For TA and DIC plots, dark gray triangles represent BATS seawater DIC and TA, pink circles represent reef seawater DIC, and purple circles represent reef seawater TA. For the coral growth rate plots, purple circles represent *P. astreoides* and green triangles represent *D. labyrinthiformis* each with  $\pm 1$  SD. Gray circles with the blue axis represent calculated NEC  $\pm$  uncertainty.

### Evaluating the relative importance of environmental controls on calcification

In an attempt to evaluate the relative importance and the interactive effects between environmental variables on colony and reef-scale calcification rates (Fig. 1), an SEM (Structural Equation Modeling) approach was used (23). This approach numerically solves the complex interactions between biotic and abiotic drivers of calcification to quantify both the direct and indirect effects of the measured environmental parameters on coral calcification. The model estimate for each reef driver (that is, temperature,  $pH_{sw}$ ,  $\Omega_A$ ,  $P_{CO_2}$ , light, nutrients, and chlorophyll *a*) repre-

sents the change in SDs in calcification explained for each SD increase in reef driver (Fig. 3 and table S1). Temperature was the only variable with significant model estimates for both species at both reef sites and NEC at Hog Reef and yielded the greatest per-SD change in coral calcification rates with calcification across the five models, increasing by 2.1 to 4.3 SDs for a single SD increase in temperature. Compared to temperature, the model estimates for the other environmental parameters ( $pH_{sw}$ ,  $\Omega_A$ ,  $P_{CO_2}$ , light, nutrients, and chlorophyll *a*) yielded smaller per-SD changes and failed to yield significant predictors for all five models (Fig. 3 and table S1). Notably, seawater  $pH_{sw}$  and  $\Omega_A$



**Fig. 3. Model estimates from the structural equation models.** Each model estimate represents the SD change in calcification driven by a 1 SD increase in the given environmental parameter. Statistically significant model estimates ( $P < 0.05$ ) are marked by stars.

produced significant predictors of calcification, but with pH<sub>sw</sub> yielding significant, negative predictors of calcification (that is, decreasing pH is correlated with increasing calcification; see subsequent discussion) for four of the five SEM models at approximately one-half of the main effect of temperature (Fig. 3 and table S1). Chlorophyll *a* produced small but significant predictors of coral calcification at Crescent Reef and NEC at Hog Reef, whereas light, P<sub>CO<sub>2</sub></sub>, and nutrients each failed to yield significant predictors of calcification for more than one of the SEM models (Fig. 3 and table S1).

Because of the narrow range of Ω<sub>A</sub> seasonal variability and highly coupled seawater pH and temperature observed during the in situ study, we performed an additional mesocosm experiment to further explore the effects of seawater carbonate chemistry on coral calcification. This mesocosm experiment was run in parallel with the in situ incubations for 3 months of the full 2-year in situ experiment using the same coral species and at three different seawater pH<sub>sw</sub> conditions (mean pH<sub>sw</sub> ± SD = 8.0 ± 0.1, 7.8 ± 0.1, and 7.6 ± 0.1). Linear models for the calcification rates of *P. astreoides* and *D. labyrinthiformis* failed to yield significant correlations between coral calcification rates and reduced seawater pH<sub>sw</sub> (7.6 ± 0.1) or Ω<sub>A</sub> (1.5 ± 0.5) relative to ambient conditions, when food was available (table S3). This finding adds to the growing literature finding that heterotrophy confers resistance to coral calcification under acidified conditions (7, 14). In contrast, coral calcification rates during the mesocosm experiment appeared more strongly correlated with temperature, lending support to the findings from the 2-year in situ experiment that seasonal calcification rates of *P. astreoides* and *D. labyrinthiformis* from Bermuda are more strongly controlled by seawater temperature when adequate nutrition is available (14).

### Predicting future coral reef calcification

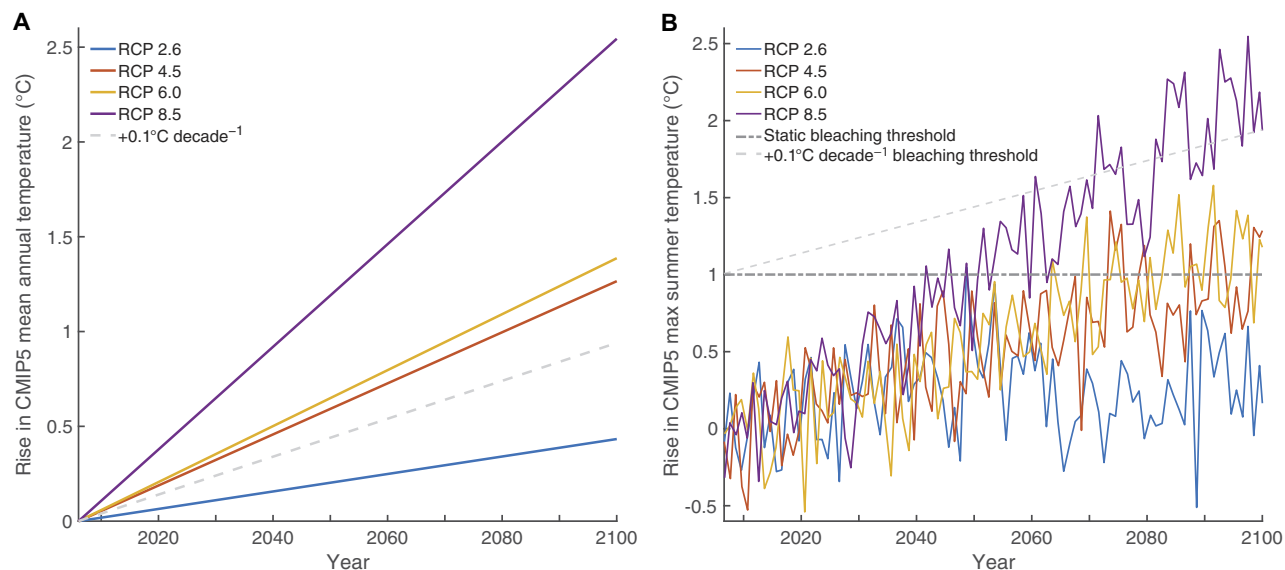
To evaluate the effect of future warming on Bermudan *P. astreoides* and *D. labyrinthiformis* coral calcification rates, data from Coupled Model Intercomparison Project Phase 5 (CMIP5) multimodel climate simulations (24) using representative concentration pathways (RCPs) of +2.6, +4.5, +6.0, and +8.5 W m<sup>-2</sup> radiative forcing relative to preindustrial levels (25) were used to simulate sea surface temperature (SST) warming

rates over the 21st century in Bermuda (Fig. 4). Linear regressions of the CMIP5 model SST predictions relative to a +0.1°C decade<sup>-1</sup> rate of warming previously observed to increase calcification in high-latitude *Porites* corals (26) show that only the most conservative emissions pathway, RCP 2.6, yields a rate of warming (+0.05°C decade<sup>-1</sup>) less than +0.1°C decade<sup>-1</sup>, suggesting that coral calcification rates could continue to increase in Bermuda under this emissions pathway (Fig. 4A). The +0.4°C end-of-21st-century seawater warming predicted for Bermuda under RCP 2.6 was combined with linear calcification responses to temperature for *P. astreoides*, *D. labyrinthiformis*, and Hog Reef NEC from the 2-year in situ study to predict potential changes in calcification rates under this reduced emissions pathway. Assuming that these relationships remain fixed over the coming century, no thermal optima are exceeded, and other environmental controls remain constant, the models suggest that *P. astreoides* and *D. labyrinthiformis* could increase by ~2 to 4% and Hog Reef NEC could increase by ~6%. The onset of coral bleaching typically occurs when warming equates to a degree heating month (DHM), wherein monthly mean SSTs exceed the maximum monthly mean climatology by 1°C with more extreme bleaching and coral mortality occurring for an annual accumulation of two DHMs (27). The CMIP5 maximum summer temperatures relative to the time period 2006–2016 show that RCP 2.6 is the only emissions scenario in which Bermudan corals are predicted to escape regular, severe coral bleaching by the end of the century on the basis of the static 1°C DHM bleaching threshold (Fig. 4B).

### DISCUSSION

The positive effects of temperature, Ω<sub>A</sub>, and chlorophyll *a* (assuming this serves as a proxy for available coral nutrition) on coral calcification rates are consistent with previous laboratory and field studies (2, 6, 7, 10–13), whereas the lack of correlation with light and the predicted positive effects owing to decreasing pH are inconsistent with anticipated results (2, 10, 13, 28). However, one has to remain circumspect about these results because the SEM is not able to elucidate functional relationships and is most likely unable to decouple the dominant





**Fig. 4. Climate model projections for Bermuda.** Monthly CMIP5 model SST data were aggregated for Bermuda under RCP emissions pathways. **(A)** Mean rise in monthly SSTs predicted for Bermuda until 2100. The dashed line represents the  $+0.1^{\circ}\text{C decade}^{-1}$  rate of potential coral acclimatization based on increasing calcification rates observed in 20th-century East Indian Ocean *Porites* coral cores (25). **(B)** Rise in maximum summer temperatures for each RCP emissions pathway relative to the  $1^{\circ}\text{C}$  coral bleaching threshold above maximum summer 2006–2016 model climatology assuming no acclimatization (Static Bleaching Threshold) and assuming a  $+0.1^{\circ}\text{C decade}^{-1}$  acclimatization rate.

effects of temperature on calcification rates and  $\text{pH}/\Omega_{\text{A}}$  from the potentially subtler effects of these carbonate chemistry parameters on calcification (9). Also, time lags between determinant properties and response variables, such as light intensity and calcification, could muddle the predictive capacity of different variables.

The temperature-induced control and seasonal increase of *P. astreoides* and *D. labyrinthiformis* coral calcification rates observed here partly agree with previous laboratory experiments and field observations, which suggest that calcification has a parabolic response to increasing seawater temperature (2, 12). In contrast, the present NEC and in situ calcification data appear to increase across the full range of seasonal temperatures, thereby suggesting that the thermal optima of the parabolic temperature–calcification response curves have not been exceeded during the 2-year study (2). Note that the in situ calcification responses to temperature may in part be affected by the coarse temporal resolution of calcification measurements (2- to 3-month skeletal growth intervals) relative to the time scales for summer seawater temperature maxima but that NEC measurements in this study reflect temperature responses over the much shorter multiday seawater residence times at Hog Reef (19). Irrespective of this, the temperature correlations in this study suggest that peak summer seawater temperatures are not limiting calcification via thermal stress and, instead, that calcification rates are more strongly limited by cooler winter seawater temperatures (15). This implies that Bermudan *P. astreoides* and *D. labyrinthiformis* coral calcification rates may increase with gradual ocean warming as has been suggested based on the analysis of interannual calcification rates from coral cores for other high-latitude coral reef environments (2). For example, a  $+0.1^{\circ}\text{C decade}^{-1}$  rate of warming yielded a 23.7% increase in calcification rates for high-latitude ( $28^{\circ}\text{C}$ ) *Porites* corals in the Houtman Abrolhos Islands off the coast of Western Australia over a 110-year period (26). Conversely, the slopes of the calcification responses to temperature in this study suggest only a ~2 to 4% increase in coral calcification rates due to the  $+0.05^{\circ}\text{C decade}^{-1}$  warming in Bermuda predicted for RCP 2.6. This potential increase is much lower than the +23.7% previously observed by Cooper *et al.* (26), suggesting that fur-

ther research should be conducted to understand the mechanisms of coral calcification responses owing to gradual ocean warming (that is,  $\leq +0.1^{\circ}\text{C decade}^{-1}$ ) and to bridge the varying insights gained from conducting laboratory experiments, in situ growth experiments, and inter-annual coral coring studies.

However, the benefits of gradual future warming must be weighed against the reduced calcification and potential mortality owing to coral bleaching (2, 10–12). For example, note that, following the coring study of Cooper *et al.* (26), an anomalous  $+5^{\circ}\text{C}$  heat wave in 2010–2011 caused coral bleaching in the Houtman Abrolhos Islands, with  $22 \pm 2.7\%$  (mean  $\pm$  SE) of corals bleached and a corresponding  $11.3 \pm 6.9\%$  (mean  $\pm$  SE) decline in overall coral cover (29). In Bermuda, there are no recorded widespread bleaching-induced mortalities, reductions in coral cover, or changes in coral community composition, but there have been some observations of coral diseases and periodic mild coral bleaching since 1988 ( $\leq 20\%$  of some coral species bleached) followed by subsequent post-bleaching recovery (15). It remains unclear whether the absence of extreme bleaching events in Bermuda is due to Bermudan reefs thus far escaping intense thermal stress events or resisting thermal stress through adaptation or acclimation [for example, see discussion of Great Barrier Reef bleaching patterns in the study of Hughes *et al.* (30)]. On the basis of the  $1^{\circ}\text{C}$  DHM static bleaching threshold (27) and the CMIP5 climate data for Bermuda, no frequent, severe bleaching is predicted for Bermudan *P. astreoides* and *D. labyrinthiformis* corals under RCP 2.6 (Fig. 4B). However, given the recent history of mild coral bleaching events ( $\leq 20\%$  of some coral species bleached) in Bermuda (15), it is likely that mild bleaching may continue in the 21st century unless Bermudan corals adapt or acclimate to these mild thermal stress events. Nonetheless, the gradual warming and absence of predicted 21st-century frequent, severe coral bleaching under RCP 2.6 in this analysis support the hypothesis that Bermudan *P. astreoides* and *D. labyrinthiformis* calcification rates could increase under RCP 2.6 throughout the 21st century with continued adequate nutrition.

However, a recent integrated assessment model utilizing recent estimates for fossil fuel resources found a 100% likelihood of exceeding

warming provided by RCP 2.6, 92% for RCP 4.5, 42% for RCP 6.0, and 12% for RCP 8.5 (31). The Paris Agreement came into effect in November 2016 following this likelihood analysis, providing an alternative pathway for individual signatory countries to collectively reduce CO<sub>2</sub> emissions to limit global warming to well below 2°C above pre-industrial levels. A 150% greater commitment to reduce CO<sub>2</sub> emissions by the current signatory countries before 2030 would yield a warming equivalent to RCP 4.5, resulting in drastically improved predicted global coral reef futures (32). Although it is less likely for Bermuda coral calcification to increase under the warming provided by RCP 4.5 (Fig. 4A), it could provide a pathway for Bermudan *P. astreoides* and *D. labyrinthiformis* corals to avoid regular severe bleaching until after 2070 (Fig. 4B), assuming no increases in coral acclimatization to future warming. Yet, there is uncertainty whether Bermudan corals can acclimatize to warming at rates faster, or slower, than the +0.1°C decade<sup>-1</sup> rate from the study of Cooper *et al.* (26). However, a +0.1°C decade<sup>-1</sup> increase in bleaching threshold could enable Bermudan corals to avoid regular severe bleaching in the 21st century under the RCP 2.6, 4.5, and 6.0 emissions pathways (Fig. 4B). In the absence of regular severe bleaching, the RCP 4.5 warming scenario therefore suggests that Bermudan *P. astreoides* and *D. labyrinthiformis* corals could maintain positive calcification at least until 2070 (Fig. 4B) with acclimatization rates of +0.1°C decade<sup>-1</sup> extending this beyond the end of the 21st century (Fig. 4B). However, note that net reef calcification and the maintenance of coral reef structure additionally depend on CaCO<sub>3</sub> dissolution and bioerosion processes, which are predicted to increase under ocean acidification (33). The calcification projections for the 21st century in this study assume that available coral nutrition remains adequate to maintain the insensitivity of adult *P. astreoides* and *D. labyrinthiformis* colonies to ocean acidification [as observed in the present mesocosm experiment and in previous studies (6, 7)] and that corals are able to successfully recruit despite the potential for negative effects of ocean acidification on scleractinian coral settlement and early life stages (34). This highlights the need for reducing CO<sub>2</sub> emissions to lessen ocean acidification impacts on Bermudan coral reef CaCO<sub>3</sub> dissolution to maintain net positive reef calcification over the coming century.

Every scleractinian species comprising the reef calcification budget for Hog Reef (*D. labyrinthiformis*, *P. astreoides*, *Pseudodiploria strigosa*, *Favia fragum*, *Madracis decactis*, *Montastrea cavernosa*, and *Orbicella franksi*) (35) is categorized as having either weedy, generalist, or stress-tolerant life histories; such species are hypothesized to be better adapted to ocean warming and acidification than faster-growing architecturally complex corals with competitive life histories (for example, *Acropora*) (36). The high natural variability of the Bermudan coral reef environment documented in this study (Crescent Reef annual  $\Delta$  seawater temperature = 12.4°C) may provide elevated thermal tolerance to corals (37) and annually variable pH<sub>sw</sub> (Hog Reef annual  $\Delta$ pH<sub>sw</sub> = 0.22) may explain the lack of sensitivity of coral calcification for adequately fed corals to predicted end-of-century ocean acidification observed in the mesocosm experiments of this study. High Bermudan coral genetic variability and population connectivity with the Caribbean may additionally increase the resilience of Bermudan coral populations during and after potential future ecological disturbances (38), with the deep reefs of Bermuda providing an additional refuge for select coral species (39). The combination of these factors and legislation protecting Bermudan coral reefs from local anthropogenic stressors (40) confer additional resiliency to this stress-tolerant coral community under current and predicted end-of-century warming and acidification. Bermudan coral reefs have maintained a relatively constant stress-tolerant, weedy,

and generalist coral community composition (table S2) (35, 41) and high coral cover (40) at least since 1980, further highlighting the resilience of Bermudan coral reefs and suggesting that reef calcification has also remained constant. That same time period in the Caribbean was characterized by declining coral cover and coral community shifts from competitive, structurally complex, and fast-growing *Acropora* corals to stress-tolerant, weedy, and generalist corals (36), driving basin-wide reductions in structural complexity (42) and net coral reef calcification (43).

These findings that Caribbean coral communities are shifting toward stress-tolerant life histories resembling the current community composition of Bermudan coral reefs suggest that a Bermuda-type coral reef system is one of many potential future stable states for coral reefs in the Caribbean and elsewhere. Characterized by anomalously high coral cover (40) and net positive reef calcification (35), coral reef systems resembling those of Bermuda may provide greater ecological resilience to 21st-century climate change and the maintenance of ecosystem services that these reef systems provide to humanity. This potentially brighter than previously predicted future for Bermudan and Caribbean coral reefs depends on mitigating other local-scale stressors (for example, overfishing and impacts of increasing human populations) (40) coupled with a continued global commitment to rapidly and drastically reducing CO<sub>2</sub> emissions to lessen the impacts of ocean warming and acidification.

## MATERIALS AND METHODS

### In situ coral calcification

Twenty-four colonies each of *P. astreoides* and *D. labyrinthiformis* were collected from Hog Reef and Crescent Reef, Bermuda, for a total of 96 coral colonies and mounted onto acrylic tiles using Z-SPAR A-788 epoxy. Twelve colonies of each species were deployed at two locations on both a rim reef site, Hog Reef, and an outer lagoon reef site, Crescent Reef, over a 2-year period (September 2010 to September 2012 at Hog Reef and August 2010 to September 2012 at Crescent Reef). Buoyant weights of each colony (16) were measured in triplicate and averaged at the end of each 2- to 3-month growth interval using a correction term for seawater density at the time of measurement and subtraction of the weight of both tile and epoxy to determine calcification as the change in weight during each growth interval. Mean initial weights ( $\pm$ SD) for *P. astreoides* were 433  $\pm$  112 g and 372  $\pm$  73 g at Hog Reef and Crescent Reef, respectively, and for *D. labyrinthiformis*, the corresponding values were 345  $\pm$  109 g and 379  $\pm$  83 g.

### Carbonate chemistry

Carbonate chemistry samples for Hog Reef and Crescent Reef were collected monthly or more frequently using a 5-liter Niskin bottle at a depth of 0.5 to 1.0 m, according to best practices (44). Offshore samples were collected monthly as part of the BATS (18). Samples for DIC and TA were collected in 200-ml Kimax glass sample bottles, fixed using 100  $\mu$ l of saturated solution of HgCl<sub>2</sub>. Reef samples of TA were analyzed via closed-cell potentiometric titrations using a Versatile Instrument for the Determination of Titration Alkalinity 3S (VINDTA 3S) system, whereas BATS samples were analyzed on a VINDTA 2S (Marianda). DIC was analyzed using coulometric methods on a VINDTA 3C or infrared-based analysis on an Automated Infra Red Inorganic Carbon Analyzer (AIRICA) system (Marianda). The accuracy and precision of TA and DIC analyses were verified against certified reference material (CRM) provided by the laboratory of A. Dickson of the Scripps Institution of Oceanography. Analysis of replicate CRMs yielded a typical

accuracy and precision of  $\pm 1$  to  $2 \mu\text{mol/kg}$  for both TA and DIC. A YSI 556 Handheld Multiparameter Instrument was used to measure in situ temperature (accuracy,  $\pm 0.15^\circ\text{C}$ ) and salinity (accuracy,  $\pm 1\%$ ), and an Autosal Salinometer (accuracy,  $<0.002$ ) was preferentially used when available to measure salinity for 66% of all bottle samples at Hog Reef and Crescent Reef. The complete carbonate system parameters (that is,  $\text{pH}_{\text{sw}}$ ,  $\text{PCO}_2$ , and  $\Omega_A$ ) were calculated using the program CO2SYS for Excel (45) and MATLAB (46) using the  $K_1$  and  $K_2$  dissociation constants from the study of Mehrbach *et al.* (47) refit by Dickson and Millero (48),  $K_{\text{H}_2\text{SO}_4}$  from Dickson (49), and pH on the seawater scale. Estimated uncertainties for calculated carbonate system parameters from the study of Millero (50) were used. Seawater and atmospheric  $\text{PCO}_2$ , temperature, salinity, and atmospheric pressure were measured autonomously every 3 hours by the National Oceanic and Atmospheric Administration Pacific Marine Environmental Laboratory (PMEL) MAPCO<sub>2</sub> (Moored Autonomous  $\text{PCO}_2$ ) moorings stationed at Hog Reef and Crescent Reef. Each mooring used a Batelle Memorial Institute CO<sub>2</sub> system to measure the mole fraction of CO<sub>2</sub> and a Sea-Bird 16plus v2 to measure temperature and salinity (51).

### Net ecosystem calcification

TA for Hog Reef and BATS bottle samples was normalized to the mean measured Hog Reef salinity of  $36.59 \text{ g kg}^{-1}$ . BATS bottle samples were linearly interpolated to match the sampling times at Hog Reef for analysis of NEC in this study. NEC was calculated for Hog Reef during the 2-year study interval and represents the net balance between calcification and  $\text{CaCO}_3$  dissolution for a well-mixed water column, as per the following equation (52)

$$\text{NEC} = \frac{\rho z (\text{TA}_{\text{offshore}} - \text{TA}_{\text{reef}})}{2\tau} \quad (1)$$

where  $\rho$  is seawater density calculated from temperature, salinity, and pressure at the time of sampling using the TEOS-10 Gibbs Seawater oceanographic toolbox (53);  $z$  is the mean  $\pm$  SD water column depth of  $10.3 \pm 3.3 \text{ m}$  calculated for Hog Reef (35);  $\text{TA}_{\text{offshore}}$  is the monthly interpolated BATS salinity normalized TA bottle sample data;  $\text{TA}_{\text{reef}}$  is the approximately monthly Hog Reef salinity normalized TA bottle sample; and  $\tau$  is the mean  $\pm$  SD seawater residence time ( $2.5 \pm 0.4$  days) calculated using a multitracers approach at the nearby North Rock rim reef site (19).

### Environmental parameters

Daily averages of temperature ( $^\circ\text{C}$ ) and light (lux) for Hog Reef and Crescent Reef were obtained from averaging four onset HOBO Pendant data loggers deployed at each reef site. Hog Reef lux data were not available for the 15, 22, and 25 September 2010 and 19 October 2010 measurements of Hog Reef NEC. A linear model constructed in MATLAB between daily Hog Reef light data and fall 2010 Bermuda Weather Service hours of sunlight [ $\text{lux} \pm \text{SE} = 146 \pm 21 \times (\text{hours of sunlight}) + 547 \pm 137$ ,  $R^2 = 0.394$ ,  $P < 0.001$ , degrees of freedom (df) = 71,  $F = 46.08$ ] was used to predict in situ Hog Reef lux data for the missing values. Monthly satellite chlorophyll *a* ( $\text{mg liter}^{-1}$ ) for the reef sites was obtained by interpolating daily chlorophyll *a* measurements from the 4-km-resolution Moderate Resolution Imaging Spectroradiometer chlorophyll *a* product. Seawater nutrient samples were taken approximately monthly and according to best practices. Nutrient samples were filtered using a  $0.4\text{-}\mu\text{m}$  filter and immediately frozen in opaque plastic bottles until processing at the Woods Hole Oceanographic Institution Nutrient Analytical Facility. All samples were analyzed on a SEAL

AutoAnalyzer 3 four-channel segmented flow analyzer using approved U.S. Environmental Protection Agency methods for ammonium (method G-171-96; detection limit  $0.034 \mu\text{M}$ ), nitrite + nitrate (method G-172-96; detection limit  $0.010 \mu\text{M}$ ), silicate (method G-177-96; detection limit  $0.016 \mu\text{M}$ ), and phosphate (method G-297-03; detection limit  $0.025 \mu\text{M}$ ). Axis 1 of a principal components analysis (PCA) performed on the ammonium, nitrite + nitrate, silicate, and phosphate data was used as a bulk nutrient metric for parameterization of the SEM analysis.

### Structural equation modeling

SEM was used to numerically solve the complex interactions between biotic and abiotic drivers of calcification to partition the direct and indirect effects of environmental drivers on calcification (23). All environmental data during the in situ coral growth monitoring at Hog Reef and Crescent Reef were monthly averaged to equally weight the data across the coral buoyant weight intervals. Coral growth was measured approximately every 2 to 3 months and was interpolated by a spline function with a cubic algorithm to link the data points and calculate monthly calcification rates so that it could be directly compared with the monthly averaged environmental data. Nutrients were synthesized by means of a standardized PCA. The first principal component was used as a proxy of nutrient variability and compared with the other environmental data. Because the Hog Reef NEC is based on a mean  $\pm$  SD seawater residence time of  $2.4 \pm 0.4$  days (19), environmental parameters that were as contemporaneous as possible to the NEC data were selected to yield the most accurate environmental parameters driving NEC. Parameters sampled at a lower resolution were interpolated to match NEC dates.

Thus, the SEM was performed on monthly averaged temperature, light, salinity,  $\Omega_A$ ,  $\text{pH}_{\text{sw}}$ , DIC, TA,  $\text{PCO}_2$ , nutrients, and chlorophyll *a* drivers of *P. astreoides* and *D. labyrinthiformis* calcification rates, and approximately daily averages for Hog Reef NEC. The strength and sign of the links and quantification of the SEM were determined by simple and partial multivariate regression and Monte Carlo permutation tests (1000 replicates), whereas chi-square values were used to assess the fit of the overall path model (54). The individual path coefficients (that is, the partial regression coefficients) indicate the relationship between the causal and response variables. Significance levels for individual paths between variables were set at  $\alpha = 0.05$ . Structural equation models were run in Amos v.21 (IBM) (55).

### Mesocosm seawater acidification experiment

The mesocosm experiments were conducted on *P. astreoides* and *D. labyrinthiformis* colonies (diameter, 12 to 16 cm) collected in July 2011 using a hammer and chisel from three Bermuda rim reef sites (near  $32^\circ 26' \text{N}$ ,  $64^\circ 50' \text{W}$ ; depth, 6 to 9 m) and exposed to three different seawater pH conditions ( $\text{pH}_{\text{sw}} = 8.0 \pm 0.1$ ,  $7.8 \pm 0.1$ , and  $7.6 \pm 0.1$ ) in controlled mesocosms for 3 months. These experiments were conducted as part of a larger-scale yearlong mesocosm experiment, including a third coral species, *Madracis auretenra*, and testing for the additional effects of feeding on calcification responses to pH treatments.

After processing each colony with an initial weighing and cementing (Z-Spar A-788) onto a preweighed tagged acrylic tile, three colonies each of *P. astreoides* and *D. labyrinthiformis* were randomly assigned to each of nine experimental tanks (three replicate tanks per pH treatment). To account for possible tank effects, coral colonies were moved around within each tank once a week, and the tanks were scrubbed clean during this time. The buoyant weight technique (16) was applied approximately monthly to each coral to calculate the calcification rate for the duration of the experiment (fig. S2).



Seawater was pumped from a nearby inlet into header tanks that each fed into three replicated experimental tanks so that conditions could more closely mimic naturally fluctuating ambient reef conditions. The two reduced pH treatments were bubbled with additional CO<sub>2</sub> into the respective header tanks with P<sub>CO2</sub> controlled by rotameters (King Instrument) to simulate seawater acidification for those treatments. Each experimental tank had flow rates of 1.67 (±0.17) liter min<sup>-1</sup>, yielding a ca. 42-min turnover time. Each tank was fed three times a week with a concentrated *Artemia* solution, resulting in approximate feeding concentrations of 1.8 mg liter<sup>-1</sup> for a 2-hour, no-flow-through period after sunset to allow corals to feed. Mesh light screens were used to protect the corals from the intense light of the shallow mesocosm tanks relative to the deeper in situ reef conditions. A YSI 556 multiparameter probe was calibrated according to standard protocols and used to monitor daily temperature, salinity, dissolved oxygen, and pH (fig. S2) with quantitative carbonate chemistry monitored by water samples analyzed for DIC, TA, and salinity. Carbonate chemistry samples were collected and analyzed, as described in the “Carbonate chemistry” section.

Before statistical analysis, all data were tested for equal variance and for normality (Kolmogorov-Smirnov test) in SigmaPlot v11 and were log-transformed (mg day<sup>-1</sup> data) or arcsine square root-transformed (mg day<sup>-1</sup> cm<sup>-2</sup> data) if necessary. Mean calcification rates for each species per tank over each growth period were analyzed using two-way repeated measures analysis of variance (ANOVA) in SigmaPlot v11 with growth periods ( $n = 4$ ) and tanks within treatment ( $n = 3$ ) as fixed main effects. Significant interactions were further analyzed with multiple comparison procedures (Holm-Sidak method). Regression analysis was used to test for significant relationships between calcification rates and calculated carbon chemistry data, seawater temperature, and light.

### Climate model predictions

The CMIP5 multimodel ensemble (24) was used in the analysis for 21st-century Bermuda SST predictions. Model output was generated by the NASA Goddard Institute for Space Studies model-run GISS-E2-H (2° latitude × 2.5° longitude 40-layer atmosphere coupled with HYbrid Coordinate Ocean Model 1° latitude × 1° longitude 26-layer ocean model) (56) using RCPs +2.6, +4.5, +6.0, and +8.5 W m<sup>-2</sup> relative to preindustrial levels (25). Monthly SST data from the four closest points surrounding Bermuda were averaged to obtain mean changes in warming anticipated for Bermuda. Linear models were fit through the monthly SST data to determine a mean rate of SST increase for the 21st century under RCP 2.6 (+0.05°C decade<sup>-1</sup>,  $P = 0.059$ ,  $df = 1138$ ,  $F = 3.57$ ), RCP 4.5 (+0.13°C decade<sup>-1</sup>,  $P < 0.0001$ ,  $df = 1138$ ,  $F = 29.8$ ), RCP 6.0 (+0.15°C decade<sup>-1</sup>,  $P < 0.0001$ ,  $df = 1138$ ,  $F = 36$ ), and RCP 8.5 (+0.27°C decade<sup>-1</sup>,  $P < 0.0001$ ,  $df = 1138$ ,  $F = 120$ ). These rates were compared to the +0.1°C decade<sup>-1</sup> warming-induced calcification increases observed by Cooper *et al.* (26) for 20th-century massive *Porites* corals in the East Indian Ocean (28°S). Because the RCP 2.6 emissions scenario predicted a rate of warming less than the +0.1°C decade<sup>-1</sup> from the study of Cooper *et al.* (26), only that emissions scenario was used to predict potential increases in coral calcification rates in Bermuda. Single-variable linear calcification responses to temperature were constructed for *P. astreoides* (Hog Reef: slope = 0.0027% day<sup>-1</sup>°C<sup>-1</sup>,  $P = 0.0003$ ,  $df = 23$ ,  $F = 17.8$ ; Crescent Reef: slope = 0.0035% day<sup>-1</sup>°C<sup>-1</sup>,  $P = 0.0002$ ,  $df = 22$ ,  $F = 20.2$ ), *D. labyrinthiformis* (Hog Reef: slope = 0.0026% day<sup>-1</sup>°C<sup>-1</sup>,  $P = 0.0006$ ,  $df = 23$ ,  $F = 15.9$ ; Crescent Reef: slope = 0.0023% day<sup>-1</sup>°C<sup>-1</sup>,  $P = 0.02$ ,  $df = 22$ ,  $F = 6.1$ ), and Hog Reef NEC (slope = 0.42 kg m<sup>-2</sup> year<sup>-1</sup>°C<sup>-1</sup>,  $P < 0.001$ ,  $df = 44$ ,  $F = 50.4$ ). The slopes of these calcification responses per degree Celsius were multiplied by

the +0.4°C predicted for end-of-21st-century Bermuda from mean RCP 2.6 warming rates to estimate potential warming-induced increases in calcification. The maximum monthly SST for each year was extracted from the output of each climate model to compare to coral bleaching thresholds. Coral bleaching is predicted to occur during a DHM in which monthly maximum summer SSTs are 1°C above mean maximum summer climatology, with an annual accumulation of two DHMs being a predictor of extreme coral bleaching and mortality (27). The first 10 years of CMIP5 maximum monthly summer SST model output (2006–2016) were averaged to approximate a mean maximum summer climatology used in calculating a relative DHM for the future warming. The difference between the summer model output and this mean maximum summer climatology was plotted to determine when maximum summer SSTs are expected to exceed the DHM of 1°C with and without a +0.1°C decade<sup>-1</sup> acclimatization rate, respectively, under RCPs 2.6, 4.5, 6.0, and 8.5.

### SUPPLEMENTARY MATERIALS

Supplementary material for this article is available at <http://advances.sciencemag.org/cgi/content/full/3/11/e1701356/DC1>

fig. S1. Hog Reef and Crescent Reef environmental data for salinity, nutrients, and satellite chlorophyll *a*.

fig. S2. Mesocosm seawater acidification experiment data.

table S1. Model estimates from the structural ecosystem models.

table S2. Bermuda coral community composition.

table S3. Mesocosm seawater acidification experiment statistical summary.

### REFERENCES AND NOTES

1. J. P. G. Spurgeon, The economic valuation of coral reefs. *Mar. Pollut. Bull.* **24**, 529–536 (1992).
2. M. S. Pratchett, K. D. Anderson, M. O. Hoogenboom, E. Widman, A. H. Baird, J. M. Pandolfi, P. J. Edmunds, J. M. Lough, Spatial, temporal and taxonomic variation in coral growth—Implications for the structure and function of coral reef ecosystems. *Oceanogr. Mar. Biol.* **53**, 215–296 (2015).
3. N. Knowlton, R. E. Brainard, R. Fisher, M. Moews, L. Plaisance, M. J. Caley, in *Life in the World's Oceans: Diversity, Distribution, and Abundance*, A. D. McIntyre, Ed. (Wiley-Blackwell, 2010), pp. 65–78.
4. G. D. Stanley Jr., Early history of scleractinian corals and its geological consequences. *Geology* **9**, 507–511 (1981).
5. B. G. Hatcher, Coral reef primary productivity: A beggar's banquet. *Trends Ecol. Evol.* **3**, 106–111 (1988).
6. K. D. Castillo, J. B. Ries, J. F. Bruno, I. T. Westfield, The reef-building coral *Siderastrea siderea* exhibits parabolic responses to ocean acidification and warming. *Proc. Biol. Sci.* **281**, 20141856 (2014).
7. E. K. Towle, I. C. Enoch, C. Langdon, Threatened Caribbean coral is able to mitigate the adverse effects of ocean acidification on calcification by increasing feeding rate. *PLOS ONE* **10**, e0123394 (2015).
8. A. McMahon, I. R. Santos, T. Cyronak, B. D. Eyre, Hysteresis between coral reef calcification and the seawater aragonite saturation state. *Geophys. Res. Lett.* **40**, 4675–4679 (2013).
9. A. Venti, A. Andersson, C. Langdon, Multiple driving factors explain spatial and temporal variability in coral calcification rates on the Bermuda platform. *Coral Reefs* **33**, 979–997 (2014).
10. J. M. Lough, N. E. Cantin, J. A. Benthuyens, T. F. Cooper, Environmental drivers of growth in massive *Porites* corals over 16 degrees of latitude along Australia's northwest shelf. *Limnol. Oceanogr.* **61**, 684–700 (2016).
11. J. M. Lough, N. E. Cantin, Perspectives on massive coral growth rates in a changing ocean. *Biol. Bull.* **226**, 187–202 (2014).
12. P. L. Jokiel, S. L. Coles, Effects of temperature on the mortality and growth of Hawaiian reef corals. *Mar. Biol.* **43**, 201–208 (1977).
13. K. Schneider, J. Erez, The effect of carbonate chemistry on calcification and photosynthesis in the hermatypic coral *Acropora eurystroma*. *Limnol. Oceanogr.* **51**, 1284–1293 (2006).
14. K. L. Yeakel, A. J. Andersson, N. R. Bates, T. J. Noyes, A. Collins, R. Garley, Shifts in coral reef biogeochemistry and resulting acidification linked to offshore productivity. *Proc. Natl. Acad. Sci. U.S.A.* **112**, 14512–14517 (2015).
15. S. R. Smith, S. Sarkis, T. J. T. Murdoch, E. Weil, A. Croquer, N. R. Bates, R. J. Johnson, S. de Putron, A. J. Andersson, Threats to coral reefs of Bermuda, in *Coral Reefs of the United*



- Kingdom Overseas Territories*, C. Sheppard, Ed. (Springer Science and Business Media, 2013), vol. 4, pp. 173–188.
16. P. L. Jokiel, J. E. Maragos, L. Franzisket, Coral growth: Buoyant weight technique, in *Coral Reefs: Research Methods*, D. R. Stoddard, R. E. Johannes, Eds. (UNESCO, 1978), pp. 529–541.
  17. S. V. Smith, Carbon dioxide dynamics: A record of organic carbon production, respiration, and calcification in the Eniwetok reef flat community. *Limnol. Oceanogr.* **18**, 106–120 (1973).
  18. N. R. Bates, M. H. P. Best, K. Neely, R. Garley, A. G. Dickson, R. J. Johnson, Detecting anthropogenic carbon dioxide uptake and ocean acidification in the North Atlantic Ocean. *Biogeosciences* **9**, 2509–2522 (2012).
  19. A. Venti, D. Kadko, A. J. Andersson, C. Langdon, N. R. Bates, A multi-tracer model approach to estimate reef water residence times. *Limnol. Oceanogr. Methods* **10**, 1078–1095 (2012).
  20. R. J. Jones, Bermuda Institute of Ocean Sciences (BIOS), Marine Environmental Program (MEP) Annual Report (2006–2007), submitted to the Bermuda Government Department of Environmental Protection, Ministry of the Environment (2006).
  21. N. R. Bates, A. Amat, A. J. Andersson, Feedbacks and responses of coral calcification on the Bermuda reef system to seasonal changes in biological processes and ocean acidification. *Biogeosciences* **7**, 2509–2530 (2010).
  22. H. C. Barkley, A. L. Cohen, D. C. McCorkle, J. Golbuu, Mechanisms and thresholds for pH tolerance in Palau corals. *J. Exp. Mar. Biol. Ecol.* **489**, 7–14 (2017).
  23. J. B. Grace, *Structural Equation Modeling and Natural Systems* (Cambridge Univ. Press, 2006).
  24. K. E. Taylor, R. J. Stouffer, G. A. Meehl, An overview of CMIP5 and the experiment design. *Bull. Am. Meteorol. Soc.* **93**, 485–498 (2012).
  25. R. H. Moss, J. A. Edmonds, K. A. Hibbard, M. R. Manning, S. K. Rose, D. P. van Vuuren, T. R. Carter, S. Emori, M. Kainuma, T. Kram, G. A. Meehl, J. F. B. Mitchell, N. Nakicenovic, K. Riahi, S. J. Smith, R. J. Stouffer, A. M. Thomson, J. P. Weyant, T. J. Wilbanks, The next generation of scenarios for climate change research and assessment. *Nature* **463**, 747–756 (2010).
  26. T. Cooper, R. A. O'Leary, J. M. Lough, Growth of Western Australian corals in the anthropocene. *Science* **335**, 593–596 (2012).
  27. S. D. Donner, W. J. Skirving, C. M. Little, M. Oppenheimer, O. Hoegh-Gulberg, Global assessment of coral bleaching and required rates of adaptation under climate change. *Glob. Chang. Biol.* **11**, 2251–2265 (2005).
  28. S. V. Smith, F. Pesret, Processes of carbon dioxide flux in the Fanning Island lagoon. *Pac. Sci.* **28**, 225–245 (1974).
  29. J. A. Y. Moore, L. M. Bellchambers, M. R. Depczynski, R. D. Evans, S. N. Evans, S. N. Field, K. J. Friedman, J. P. Gilmour, T. H. Holmes, R. Middlebrook, B. T. Radford, T. Ridgway, G. Shedrawi, H. Taylor, D. P. Thomson, S. K. Wilson, Unprecedented mass bleaching and loss of coral across 12° of latitude in Western Australia in 2010–11. *PLOS ONE* **7**, e51807 (2012).
  30. T. P. Hughes, J. T. Kerry, M. Álvarez-Noriega, J. G. Álvarez-Romero, K. D. Anderson, A. H. Baird, R. C. Babcock, M. Beger, D. R. Bellwood, R. Berkelmans, T. C. Bridge, I. R. Butler, M. Byrne, N. E. Cantin, S. Comeau, S. R. Connolly, G. S. Cumming, S. J. Dalton, G. Diaz-Pulido, C. M. Eakin, W. F. Figueira, J. P. Gilmour, H. B. Harrison, S. F. Heron, A. S. Hoey, J.-P. A. Hobbs, M. O. Hoogenboom, E. V. Kennedy, C.-K. Kuo, J. M. Lough, R. J. Lowe, G. Liu, M. T. McCulloch, H. A. Malcolm, M. J. McWilliam, J. M. Pandolfi, R. J. Pears, M. S. Pratchett, V. Schoepf, T. Simpson, W. J. Skirving, B. Sommer, G. Torda, D. R. Wachenfeld, B. L. Willis, S. K. Wilson, Global warming and recurrent mass bleaching of corals. *Nature* **543**, 373–377 (2017).
  31. I. Capellán-Pérez, I. Arto, J. M. Polanco-Martínez, M. González-Eguino, M. B. Neumann, Likelihood of climate change pathways under uncertainty on fossil fuel resource availability. *Energy Environ. Sci.* **9**, 2482–2496 (2016).
  32. R. van Hooidonk, J. Maynard, J. Tamelander, J. Gove, G. Ahmadi, L. Raymundo, G. Williams, S. F. Heron, S. Planes, Local-scale projections of coral reef futures and implications of the Paris Agreement. *Sci. Rep.* **6**, 39666 (2016).
  33. B. D. Eyre, A. J. Andersson, T. Cyronak, Benthic coral reef calcium carbonate dissolution in an acidifying ocean. *Nat. Clim. Chang.* **4**, 969–976 (2014).
  34. R. Albright, C. Langdon, Ocean acidification impacts multiple early life history processes of the Caribbean coral *Porites astreoides*. *Glob. Chang. Biol.* **17**, 2478–2487 (2011).
  35. T. A. Courtney, A. J. Andersson, N. R. Bates, A. Collins, T. Cyronak, S. J. de Putron, B. D. Eyre, R. Garley, E. J. Hochberg, R. Johnson, S. Musielewicz, T. J. Noyes, C. L. Sabine, A. J. Sutton, J. Tonicin, A. Tribollet, Comparing chemistry and census-based estimates of net ecosystem calcification on a rim reef in Bermuda. *Front. Mar. Sci.* **3**, 181 (2016).
  36. E. S. Darling, L. Alvarez-Filip, T. A. Oliver, T. R. McClanahan, I. M. Côté, Evaluating life-history strategies of reef corals from species traits. *Ecol. Lett.* **15**, 1378–1386 (2012).
  37. T. A. Oliver, S. R. Palumbi, Do fluctuating temperature environments elevate coral thermal tolerance? *Coral Reefs* **30**, 429–440 (2011).
  38. J. M. Locke, K. A. Coates, J. P. Bilewicz, L. P. Holland, J. M. Pitt, S. R. Smith, H. G. Trapido-Rosenthal, Biogeography, biodiversity and connectivity of Bermuda's coral reefs, in *Coral Reefs of the United Kingdom Overseas Territories*, C. Sheppard, Ed. (Springer Science and Business Media, 2013), pp. 153–172.
  39. Pim Bongaerts, Cynthia Riginos, Ramona Brunner, Norbert Englebert, Struan R. Smith, Ove Hoegh-Guldberg, Deep reefs are not universal refuges: Reseeding potential varies among coral species. *Sci. Adv.* **3**, e1602373 (2017).
  40. J. B. C. Jackson, M. K. Donovan, K. L. Cramer, V. V. Lam, *Status and Trends of Caribbean Coral Reefs: 1970–2012* (Global Coral Reef Monitoring Network, IUCN, 2014).
  41. R. E. Dodge, A. Logan, A. Antonius, Quantitative reef assessment studies in Bermuda: A comparison of methods and preliminary results. *Bull. Mar. Sci.* **32**, 745–760 (1982).
  42. L. Alvarez-Filip, N. K. Dulvy, J. A. Gill, I. M. Côté, A. R. Watkinson, Flattening of Caribbean coral reefs: Region-wide declines in architectural complexity. *Proc. Biol. Sci.* **276**, 3019–3025 (2009).
  43. C. T. Perry, G. N. Murphy, P. S. Kench, S. G. Smithers, E. N. Edinger, R. S. Steneck, P. J. Mumby, Caribbean-wide decline in carbonate production threatens coral reef growth. *Nat. Commun.* **4**, 1402 (2013).
  44. A. G. Dickson, C. L. Sabine, J. R. Christian, *Guide to Best Practices for Ocean CO<sub>2</sub> Measurements* (PICES Special Publication, 2007), vol. 3, 191 pp.
  45. E. Lewis, D. Wallace, *Program Developed for CO<sub>2</sub> System Calculations: Carbon Dioxide Information Analysis Center, Oak Ridge National Laboratory, ORNL/CDIAC-105* (US DOE, 1998).
  46. S. Van Heuven, D. Pierrot, J. W. B. Rae, E. Lewis, D. W. R. Wallace, MATLAB program developed for CO<sub>2</sub> system calculations, ORNL/CDIAC-105b, Carbon Dioxide Inf. Anal. Cent., Oak Ridge Natl. Lab., US DOE, Oak Ridge, TN (2011).
  47. C. Mehrbach, C. H. Culberson, J. E. Hawley, R. M. Pytkowicz, Measurement of the apparent dissociation constants of carbonic acid in seawater at atmospheric pressure. *Limnol. Oceanogr.* **18**, 897–907 (1973).
  48. A. G. Dickson, F. J. Millero, A comparison of the equilibrium constants for the dissociation of carbonic acid in seawater media. *Deep Sea Res.* **34**, 1733–1743 (1987).
  49. A. G. Dickson, Thermodynamics of the dissociation of boric acid in synthetic seawater from 273.15 to 318.15 K. *Deep Sea Res.* **37**, 755–766 (1990).
  50. F. J. Millero, The marine inorganic carbon cycle. *Chem. Rev.* **107**, 308–341 (2007).
  51. A. J. Sutton, C. L. Sabine, S. Maenner-Jones, N. Lawrence-Slavas, C. Meinig, R. A. Feely, J. T. Mathis, S. Musielewicz, R. Bott, P. D. McLain, H. J. Fought, A. Kozyr, A high-frequency atmospheric and seawater pCO<sub>2</sub> data set from 14 open-ocean sites using a moored autonomous system. *Earth Syst. Sci. Data* **6**, 353–366 (2014).
  52. C. Langdon, J.-P. Gattuso, A. Andersson, in *Guide to Best Practices for Ocean Acidification Research and Data Reporting*, U. Riebesell, V. Fabry, L. Hansson, J. Gattuso, Eds. (Office for Official Publications of the European Communities, 2010), pp. 213–232.
  53. T. J. McDougall, P. M. Barker, *Getting Started with TEOS-10 and the Gibbs Seawater (GSW) Oceanographic Toolbox* (SCOR/IAPSO WG 127, 2011), 28 pp.
  54. C. Alsterberg, J. S. Eklöf, L. Gamfeldt, J. N. Havenhand, K. Sundbäck, Consumers mediate the effects of experimental ocean acidification and warming on primary producers. *Proc. Natl. Acad. Sci. U.S.A.* **110**, 8603–8608 (2013).
  55. J. L. Arbuckle. Amos (Version 21.0) [Computer Program]. Chicago: IBM SPSS (2014).
  56. G. A. Schmidt, M. Kelley, L. Nazarenko, R. Ruedy, G. L. Russell, I. Aleinov, M. Bauer, S. E. Bauer, M. K. Bhat, R. Bleck, V. Canuto, Y.-H. Chen, Y. Cheng, T. L. Clune, A. Del Genio, R. de Fainchtein, G. Faluvegi, J. E. Hansen, R. J. Healy, N. Y. Kiang, D. Koch, A. A. Lacis, A. N. LeGrande, J. Lerner, K. K. Lo, E. E. Matthews, S. Menon, R. L. Miller, V. Oinas, A. O. Olosio, J. P. Perlwitz, M. J. Puma, W. M. Putman, D. Rind, A. Romanou, M. Sato, D. T. Shindell, S. Sun, R. A. Syed, N. Tausnev, K. Tsigaridis, N. Unger, A. Voulgarakis, M.-S. Yao, J. Zhang, Configuration and assessment of the GISS ModelE2 contributions to the CMIP5 archive. *J. Adv. Model. Earth Syst.* **6**, 141–184 (2014).

**Acknowledgments:** This project was made possible by the countless contributions by researchers, staff, volunteers, and interns at the Bermuda Institute of Ocean Sciences. We acknowledge the World Climate Research Programme's Working Group on Coupled Modelling for Coupled Model Intercomparison Project and thank the NASA Goddard Institute for Space Studies for producing and making their CMIP5 model output available. We thank F. Melin for extracting and providing the satellite chlorophyll *a* data. Additionally, we thank two anonymous reviewers for improving this manuscript with their constructive comments and feedback. This is PMEL contribution number 4705. **Funding:** This work was funded by National Science Foundation (NSF) OCE 09-28406 (to A.J.A., N.R.B., and S.J.d.P.), NSF OCE 14-16518 (to A.J.A. and R.J.), and the NSF Graduate Research Fellowship Program (to T.A.C.). **Author contributions:** A.J.A., N.R.B., and S.J.d.P. designed the study. A.J.A., N.R.B., S.J.d.P., A.C., R.G., R.J., T.J.N., and C.L.S. conducted measurements and/or contributed data. T.A.C., A.J.A., M.L., and J.-C.M. analyzed the data and results. T.A.C. and A.J.A. wrote the manuscript with input from all coauthors. **Competing interests:** The authors declare that they have no competing interests. **Data availability:** All data needed to evaluate the conclusions in the paper are present in the paper and/or the Supplementary Materials. Additional data related to this paper may be requested from the authors. The data presented in this manuscript will be made available in the Biological and Chemical Oceanography Data Management Office database under the BEACON project (<http://bco-dmo.org/project/2190>).

Submitted 26 April 2017

Accepted 17 October 2017

Published 8 November 2017

10.1126/sciadv.1701356

**Citation:** T. A. Courtney, M. Lebrato, N. R. Bates, A. Collins, S. J. de Putron, R. Garley, R. Johnson, J.-C. Molinero, T. J. Noyes, C. L. Sabine, A. J. Andersson, Environmental controls on modern scleractinian coral and reef-scale calcification. *Sci. Adv.* **3**, e1701356 (2017).

## Environmental controls on modern scleractinian coral and reef-scale calcification

Travis A. Courtney, Mario Lebrato, Nicholas R. Bates, Andrew Collins, Samantha J. de Putron, Rebecca Garley, Rod Johnson, Juan-Carlos Molinero, Timothy J. Noyes, Christopher L. Sabine and Andreas J. Andersson

*Sci Adv* 3 (11), e1701356.  
DOI: 10.1126/sciadv.1701356

### ARTICLE TOOLS

<http://advances.sciencemag.org/content/3/11/e1701356>

### SUPPLEMENTARY MATERIALS

<http://advances.sciencemag.org/content/suppl/2017/11/06/3.11.e1701356.DC1>

### REFERENCES

This article cites 43 articles, 7 of which you can access for free  
<http://advances.sciencemag.org/content/3/11/e1701356#BIBL>

### PERMISSIONS

<http://www.sciencemag.org/help/reprints-and-permissions>

Use of this article is subject to the [Terms of Service](#)

---

*Science Advances* (ISSN 2375-2548) is published by the American Association for the Advancement of Science, 1200 New York Avenue NW, Washington, DC 20005. 2017 © The Authors, some rights reserved; exclusive licensee American Association for the Advancement of Science. No claim to original U.S. Government Works. The title *Science Advances* is a registered trademark of AAAS.